

## WHAT HAVE WE LEARNT ABOUT MEDITERRANEAN CATCHMENT HYDROLOGY? 30 YEARS OBSERVING HYDROLOGICAL PROCESSES IN THE VALLCEBRE RESEARCH CATCHMENTS

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**ABSTRACT.** *This paper presents the main results obtained from the study of hydrological processes in the Vallcebre Research Catchments since 1988. Distributed hydrometric measurements, environmental tracers and hydrological modelling were used to understand Mediterranean catchment behaviour and to provide new data to help assess the global change effects on these catchments' water resources. Thirty years of hydrological processes observation in the Vallcebre Research Catchments have increased understanding not only of their hydrological response, but also of the main hydrological and erosion processes characteristic of Mediterranean mountain catchments. This paper briefly summarises the main results obtained since 1988 on ecohydrological processes, hydrological response, runoff generation processes, erosion and sediment transport. Some of the main findings from this research are (i) the importance of temporal variability in precipitation to determine the hydrological processes; (ii) the paramount role played by forest cover in reducing soil water content; (iii) the marked influence of antecedent wetness conditions on runoff generation that determine different runoff responses; (v) the dominant contribution of pre-existing water during floods; (vi) the importance of freezing-thawing processes in badland areas on erosion and the role of summer convective storms in controlling sediment transport.*

**¿Qué hemos aprendido sobre la hidrología de cuencas Mediterráneas? 30 años observando los procesos hidrológicos en las cuencas de investigación de Vallcebre**

**RESUMEN.** *Este trabajo presenta los principales resultados sobre el estudio de los procesos hidrológicos obtenidos en las cuencas de investigación de Vallcebre desde 1988. Se han utilizado aproximaciones metodológicas complementarias, tales como mediciones hidrométricas distribuidas, trazadores ambientales y modelización hidrológica con la finalidad de mejorar la comprensión de la respuesta de las cuencas mediterráneas y proporcionar nuevos conocimientos válidos para evaluar los efectos*

*de cambio global en los recursos hídricos de estas cuencas. Treinta años de observaciones de los procesos hidrológicos en las cuencas de investigación de Vallcebre han permitido comprender no sólo su respuesta hidrológica, sino también los principales procesos hidrológicos y de erosión propios de las cuencas mediterráneas. Este artículo resume los principales resultados obtenidos desde 1988 sobre los procesos ecohidrológicos, la respuesta hidrológica, los procesos de generación de escorrentía, la erosión y el transporte de sedimentos. Algunos de los principales resultados de estas investigaciones son (i) la crucial importancia de la variabilidad temporal de las precipitaciones en los procesos hidrológicos; (ii) el papel fundamental de la cubierta forestal en la reducción del contenido hídrico del suelo; (iii) la marcada influencia de las condiciones de humedad antecedentes en la generación de escorrentía y su efecto en las diferentes respuestas hidrológicas; v) la contribución dominante de las aguas preexistentes durante las crecidas; vi) la importancia que tienen sobre la erosión los procesos de hielo-deshielo en las áreas de badlands y el papel de las tormentas convectivas de verano en el transporte de sedimentos.*

**Key words:** hydrological processes, environmental tracers, Vallcebre research catchments, Mediterranean mountain area.

**Palabras clave:** procesos hidrológicos, trazadores ambientales, cuencas de investigación de Vallcebre, área de montaña mediterránea.

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## 1. Introduction

The study and modelling of hydrological processes in the Vallcebre research catchments started about 30 years ago with the aim of improving the understanding of hydrological processes that control the seasonality of the response in Mediterranean catchments. The research area was selected in 1988 to analyse the hydrological consequences of land abandonment, as well as the hydrological and sediment yield behaviour of badlands areas. The approach used combines various complementary methodologies, such as distributed hydrometric measurements, environmental tracers and hydrological modelling, and provides new insights into the effects of global change on water resources. This paper briefly summarises the main results obtained since 1988 on the spatio-temporal dynamics of precipitation, rainfall partitioning, forest transpiration and ecohydrological processes, soil water and groundwater dynamics, hydrological response and runoff generation processes, erosion and sediment transport, as well as some results derived from modelling at several spatial and temporal scales.

Our continuous concern for the field study of hydrological processes in the Vallcebre research catchments is our grain of sand to add to recent opinion papers (e.g. Burt and

McDonnell, 2015) advocating the continuity of field studies in catchment hydrology. This continuity is especially important at a time when it is more efficient to produce papers on modelling than on field studies (Burt and McDonnell, 2015; Beven, 2016). We are indeed convinced that, as indicated by Beven (2016), measurement and fieldwork are unique tools for understanding environmental processes and developing new explanations of how catchments function. We also agree that there is a need not only for short-term experiments or campaigns, sometimes only designed to calibrate and validate hydrological models, but also for long-term data from research catchments to further process understanding, observe trends and test new ideas, methods or instruments (Burt and McDonnell, 2015; Tetzlaff *et al.*, 2017).

## 2. Study site

### 2.1. General setting

The Vallcebre research area is located close to Vallcebre village (Spain), at the headwaters of the Llobregat river, about 100 km North of Barcelona, in the South-eastern part of the Pyrenees (42°12'N and 1°49'E) (Fig. 1). The Llobregat River provides the main water supply for the Barcelona area.

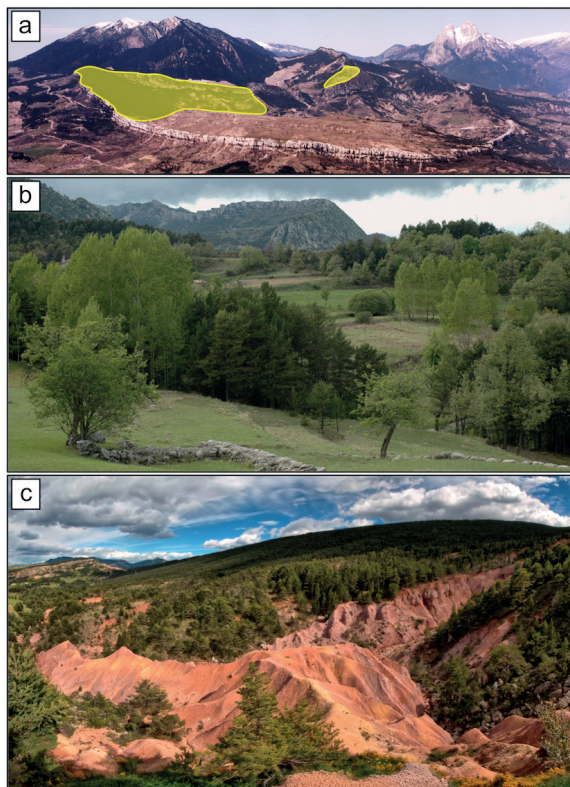


Figure 1. Map of the Vallcebre Research Catchments, showing locations of the currently monitored forest plots and of the main instruments operating in 2016.

Within the Vallcebre research area, the Surface Hydrology and Erosion group of the IDAEA\_CSIC has monitored two clusters of five catchments (Fig. 2a). The main cluster, the Cal Rodó catchment (4.17 Km<sup>2</sup>), has two sub-catchments: Can Vila (0.56 Km<sup>2</sup>) and Ca l'Isard (1.32 km<sup>2</sup>). The smallest cluster, Cal Parisa (0.32 km<sup>2</sup>) is formed by two contiguous catchments of similar size. Long-term monitoring of forest plots, covered by Scots pines or by pubescent oaks, are also part of this research facility (Fig. 1).

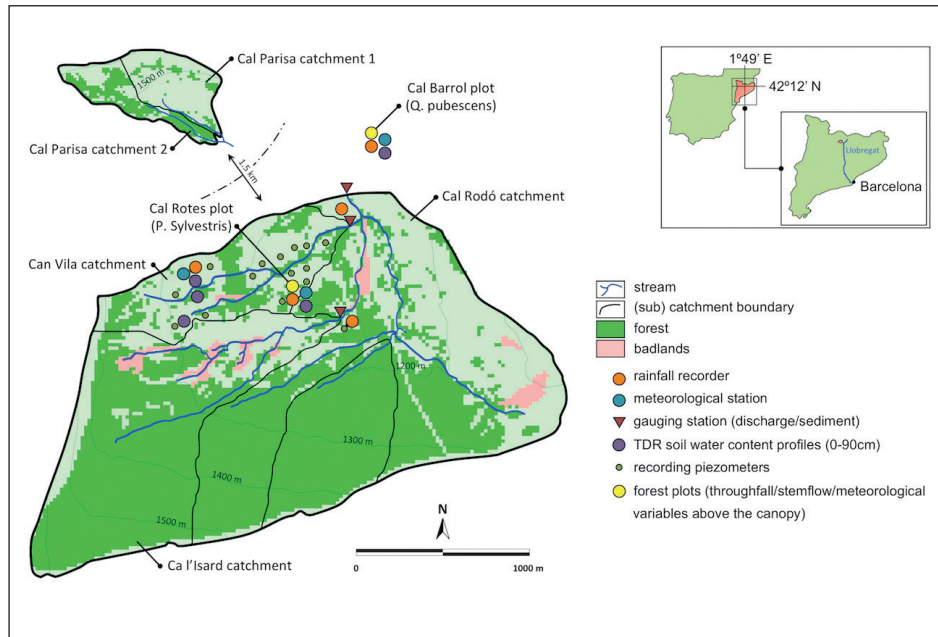


Figure 2. General view of Vallcebre Research Catchments (a). The smaller catchment on the right is the Cal Parisa catchment and the larger one on the left is the Cal Rodó catchment. The catchments are covered by pastures and forest (b) with some small badlands areas (c) close to the stream.

## 2.2. Topography

Within the Vallcebre research area, altitudes range between 1100 and 1700 m a.s.l. Slopes vary between 10% and 40% (mean value 28%) and mostly face northeast. The topographical limits of the catchments are well defined, and correspond in over half the Cal Parisa and Cal Rodó catchments to the same limestone bed folded in a near-vertical plane.

Drainage density is relatively low (2.6 km km<sup>-2</sup>) and channels are only fully active during floods. Conversely, during baseflow conditions less than half the total channel length is active. The main channels are 1 to 3 m wide and are not very deeply incised. The streambeds follow mostly a riffle-pool sequence, with the materials mainly formed

by coarse alluvium partly cemented by lime coatings. Mobile sediments are mostly fine sands and silt. There are no riparian zones in the catchments.

Before and during the 19th century most of the hillslopes were deforested and terraced for agricultural purposes. The size of the terraces depends greatly on the local gradient, but most of them are 10 to 20 m wide, often limited by a stone wall up to 2 m high. The terraced topography covers around half the area. Along with terraces, a network of ditches was also built in order to prevent soil saturation and to convey surface runoff (Llorens *et al.*, 1992; Gallart *et al.*, 1994). However, this artificial drainage network is less extensive than the natural stream, though in some places the differentiation between artificial and natural stream is not straightforward.

### 2.3. Vegetation

Climax vegetation is woodland of pubescent oak (*Quercus pubescens*) with Scots pine (*Pinus sylvestris*) on the cooler north-facing aspects. Most of the gentle hillslopes, which were deforested in the past and terraced, are now covered by mesophile grasses with hydrophile patches of *Molinia coerulea* and are used for either grass harvesting or extensive sheep grazing. Following land abandonment in the sixties, spontaneous forestation by *Pinus sylvestris* occurred (Poyatos *et al.*, 2003) and forest now covers half the area (Fig. 2b). The remainder of the area is widely covered by pasture and meadows, along with small areas of sparse scrub vegetation and bare soil.

The three pine plots studied (Cal Parisa, Cal Sort and Cal Rotes) are covered by monospecific *Pinus sylvestris* stands. These plots were selected as representative of patches of pines overgrown in marginal terraced areas that were abandoned in the second half of the 20th century. The plots are rather young stands (the oldest trees are about 70 years old) with scarce understorey (mainly *Buxus sempervirens*). Stand density varies between 1,189 and 2,400 trees ha<sup>-1</sup>, with mean diameter at breast height (DBH) between 17 and 25 cm. The *Quercus pubescens* plot (Cal Barrol) is characterized by the presence of other woody species (*Prunus avium* L., *Fraxinus excelsior* L.) and a dense understorey (*Acer campestre* L., *Buxus sempervirens* L., *Prunus spinosa* L., *Rubus spp.* and *Rosa Spp.*) Stand density is 518 trees ha<sup>-1</sup> and mean DBH about 20 cm.

### 2.4. Geology and soils

In the South half of the Cal Rodó catchment, bedrock is massive limestone of continental Garumnian facies (Aepler, 1968), with very few conspicuous karstic features. Elsewhere bedrock is formed by red clayey smectite-rich mudrocks. The soft mudstones are prone to landslides and erosion by water, leading to intensely dissected landscapes with poor vegetation cover (badlands) as a result of the combined effects of winter freezing and summer rainstorms (Regüés *et al.*, 1995). Badlands areas (Fig. 2c) occupy less than 3% of the Cal Rodó catchment area, but play a dominant role in sediment production (Gallart *et al.*, 2005).

Soil thickness varies greatly, depending on lithology, geomorphology and terracing. Limestone areas are overlain by discontinuous soils less than 40 cm deep, whereas soils on hillslopes over clayey rocks may reach a depth of 80 cm. Badlands areas have regoliths whose thickness varies throughout the year, but which rarely reach 15 cm. Finally, soil thickness ranges from less than 50 cm in the inner part of the terraces to more than 2 or 3 m in their outer part (Latron *et al.*, 2008).

Soils have a silty loam and silty clay loam texture. Bulk density increases with increasing depth, from 0.85 g cm<sup>-3</sup> in the top layer to 1.65 g cm<sup>-3</sup> at 50 cm. On the contrary, organic matter decreases with increasing depth from 15.3% to 0.3% (Rubio *et al.*, 2008). These characteristics promote high infiltration capacities in the first 20 cm below the surface (Solé *et al.*, 1992) and the dropping of hydraulic conductivity by several orders of magnitude in the deeper horizons (Haro *et al.*, 1992; Rubio, 2005), inducing the formation of shallow semi-permanent aquifers, despite the high topographic gradients. The arrangement of agricultural terraces also results in the great spatial variability of hydraulic conductivities, which tend to be much higher near the outer edges than in the internal parts. Soil cracking during summer is favoured by smectite clay content and enhances infiltration capacity, whereas soil crusting was only observed on bare badlands surfaces (Regüés *et al.*, 1995).

### 3. Field measurements, monitoring and sampling

In addition to the monitoring of meteorological variables, precipitation, discharge and sediment concentration measured at the catchment outlet for almost 3 decades, several spatially distributed state variables (soil water content and piezometric levels) and fluxes (rainfall partitioning and transpiration) have also been recorded, though for shorter periods. Moreover, since 2011 rainwater, throughfall, stemflow, runoff and several hydrological compartments (soil water, groundwater and vegetation) have been sampled intensively for isotopic determinations. Figure 3 shows some details of the monitoring set-ups (catchment and plot scales) for the measurement and sampling of hydrological variables and fluxes; and Table 1 shows the location of measurement sites and the periods of data recording and sampling.

Data on rainfall partitioning and forest transpiration were obtained at one Scots pine plot located in the Cal Parisa catchment (Cal Parisa plot), two other Scots pine plots located in the Cal Rodó catchment (Cal Sort and Cal Rotes plots) and a Quercus plot located near the Cal Rodó outlet (Cal Barrol plot). Soil water content data on the catchment scale were obtained in a set of TDR profiles at both the Cal Parisa and Can Vila catchments and, on the plot scale, from a densely monitored terrace in the Can Vila catchment. Distributed water table data were obtained mainly from the Can Vila catchment. Main results on hydrological response and runoff generation processes were obtained from the Can Vila catchment. The main reason for selecting this catchment instead of Cal Rodó was that Can Vila is totally watertight. Indeed, the geological setting of the Cal Rodó catchment has raised some questions about its hydrological limits and the issue of its water tightness. By establishing a monthly water balance between the Cal Rodó catchment and the Can Vila sub-catchment, entirely on clayey bedrock, deep

percolation through the limestone was calculated over a period of 26 months as 21% of incoming rainfall (Latron, 2003). The ratio between monthly loss through the limestone and precipitation is relatively similar during dry and wet periods and deep percolation has been found to affect essentially baseflow volumes and not storm-flow ones. Finally, erosion processes and suspended sediment transport data have been investigated mainly in the Ca l'Isard catchment, where most of the badlands areas are located. Within this catchment, an elementary badlands area of 1200 m<sup>2</sup> (El Carot) was also instrumented for 3 years.

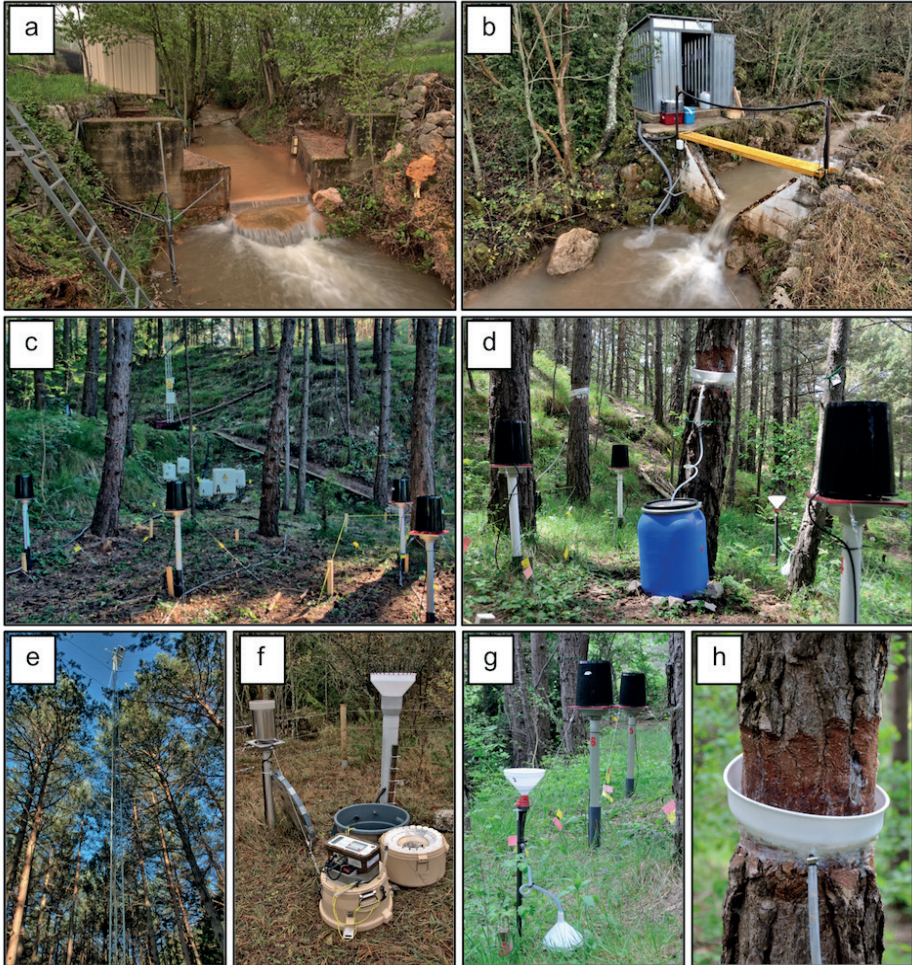


Figure 3. Gauging station of the Cal Rodó catchment (4.17 km<sup>2</sup>) (a) and of the Can Vila sub-catchment (0.56 km<sup>2</sup>) (b). Throughfall collectors, stemflow rings in the Cal Rotes plots (Scots Pine) (c, d, g, h). 18-meter meteorological tower in the Cal Rotes plot for measuring above-canopy conditions (e). Rainfall recorder and sequential rainfall sampler (5 mm rainfall increment) (f).

Table 1. Location of measurement sites (hydrological variables and fluxes) and periods of data recording and sampling in the Vallecbre Research Catchments.

Variables and fluxes	Measuring period	Temporal resolution	NMP1	Instruments	Catchment/plot
<b>MONITORING</b>					
<b>Precipitation</b>	1982–1989	Daily	1	rain gauge	Vallecbre
	1989–2004	5min	3	rain recorders	Cal Parisa
	1996–cont.	5min	2–10	rain recorders	Cal Rodó
<b>Meteorological variables<sup>2</sup></b>	1989–2001	5min	1	standard meteorological station	Cal Parisa
	1997–cont.	5min	1	standard meteorological station	Can Vila
	2003–cont.	5min	1	meteorological tower	Cal Sort, Cal Rotes ( <i>P. sylvestris</i> )
	2003–cont.	5min	1	meteorological tower	Cal Barrol ( <i>Q. pubescens</i> )
<b>Stream flow level</b>	1989–2004	5min	2	gauging stations with water level sensors	Cal Parisa
	1993–cont.	5min	3	gauging stations with water level sensors	Cal Rodó, Ca' l'Isard, Can Vila
<b>Groundwater level</b>	1994–2003	weekly	1	manually measured depth to water table	Cal Parisa
	1995–2006	20min	3	piezometers	Can Vila
	2006–cont.	10min	20	piezometers	Can Vila
<b>Soil water content</b>	1993–2011	weekly	8	TDR profile 0–80 cm (4 x 20cm probes)	Cal Parisa (3 meadows, 2 forest)
	1996–cont.	weekly	2	TDR profile 0–80 cm (4 x 20cm probes)	Can Vila (meadows)
	2006–cont.	weekly	3	TDR profile 0–80 cm (4 x 20cm probes)	Can Vila (2 meadows, 1 forest)
	2012–cont	weekly	1	TDR profile 0–90 cm (3 x 30cm probes)	Cal Rodó ( <i>P. sylvestris</i> )
	2006–cont.	5min	1	automatic TDR profile 0–90 (3x30cm probes)	Can Vila (meadows)
	2009–cont.	5min	1	automatic TDR profile 0–90 (3x30cm probes)	Can Vila (meadows)
	2012–cont.	5min	1	automatic TDR profile 0–90 (3x30cm probes)	Can Rotes ( <i>P. sylvestris</i> )
	2005–2007	20min	128	TDR probes 30cm	Can Vila (terrace meadows)
	2012–2013	20min	20	TDR probes 20cm	Cal Rotes ( <i>P. sylvestris</i> )
	2012–2013	20min	40	TDR probes 20cm	Cal Barrol ( <i>Q. pubescens</i> )
	2017–cont.	20min	68	TDR probes around a tree (10, 20 and 30cm)	Cal Rotes ( <i>P. sylvestris</i> )
<b>Throughfall</b>	1993–2003	5min	9	throughfall troughs with tipping buckets	Cal Parisa ( <i>P. sylvestris</i> )
	2004–2009	5min	6	throughfall troughs with tipping buckets	Cal Barrol ( <i>Q. pubescens</i> )
	2004–2009	5min	20	throughfall collectors with tipping buckets	Cal Barrol ( <i>Q. pubescens</i> )
	2012–2016	5min	20	throughfall collectors with tipping buckets	Cal Rotes ( <i>P. sylvestris</i> )
	2012–2016	5min	20	throughfall collectors with tipping buckets	Cal Rotes ( <i>P. sylvestris</i> )
<b>Stemflow</b>	1993–2003	5min	6	stemflow rings with tipping buckets	Cal Rotes ( <i>P. sylvestris</i> )
	2004–2009	5min	7	stemflow rings with tipping buckets	Cal Barrol ( <i>Q. pubescens</i> )
	2012–2016	5min	7	stemflow rings with tipping buckets	Cal Rotes ( <i>P. sylvestris</i> )
	2012–2016	5min	7	stemflow rings with tipping buckets	Cal Barrol ( <i>Q. pubescens</i> )
	2017–cont.	5min	2	stemflow rings separated in 4 sections with tipping buckets	Cal Rotes ( <i>P. sylvestris</i> )



Variables and fluxes	Measuring period	Temporal resolution	NMP <sup>1</sup>	Instruments	Catchment/plot
<b>Trees transpiration</b>					
	1994–2000	15min	7	trees with thermal dissipation sapflow probes	Cal Parisa ( <i>P. sylvestris</i> )
	2003–2005	15min	12	trees with thermal dissipation sapflow probes	Cal Sort ( <i>P. sylvestris</i> )
	2003–2005	15min	12	trees with thermal dissipation sapflow probes	Cal Barrol ( <i>Q. pubescens</i> )
<b>SAMPLING</b>					
<b>Water isotopes<sup>3</sup></b>					
<b>Rainfall</b>					
	2011-cont.	weekly	1-4	bulk rainfall samplers	Can Vila
	2011-2013	5mm of rainfall	1	automatic sampler	Can Vila
	2015-2016	5mm of rainfall	1-2	automatic sampler	Can Vila
	2017-cont.	5mm of rainfall	2	automatic sampler	Can Vila
<b>Stream water</b>					
	2011-cont.	weekly	1	manual sampling	Can Vila
	2011-cont.	during floods	2	automatic sampler	Can Vila
<b>Groundwater, soil water</b>					
	2011-2013	biweekly	4	groundwater manual sampling	Can Vila
	2011-2013	biweekly	2	soil water manual sampling	Can Vila
<b>Throughfall</b>					
	2015-2016	weekly	10	throughfall bulk collectors	Cal Rotes ( <i>P. sylvestris</i> )
	2015-2016	weekly	10	throughfall bulk collectors	Cal Barrol ( <i>Q. pubescens</i> )
	2015-2016	5mm of rainfall	1	automatic sampler	Cal Rotes ( <i>P. sylvestris</i> )
	2015-2017	5mm of rainfall	1	automatic sampler	Cal Barrol ( <i>Q. pubescens</i> )
	2017-cont.	weekly	10	throughfall bulk collectors	Cal Rotes ( <i>P. sylvestris</i> )
	2017-cont.	5mm of rainfall	1	automatic sampler	Cal Rotes ( <i>P. sylvestris</i> )
<b>Stemflow</b>					
	2015-2016	weekly	4	stemflow rings with collectors	Cal Rotes ( <i>P. sylvestris</i> )
	2015-2016	weekly	4	stemflow rings with collectors	Cal Barrol ( <i>Q. pubescens</i> )
	2017-cont.	weekly	3	stemflow rings with collectors	Cal Rotes ( <i>P. sylvestris</i> )
	2017-cont.	5min	1	automatic sampler	Cal Rotes ( <i>P. sylvestris</i> )
<b>Suspended sediment concentration</b>					
	1989-2004	during floods	2	automatic samplers	Cal Parisa
	1995-cont.	during floods	6	automatic samplers	Cal Rodó, Can Vila, Ca l'Isard
	1995-cont.	5min	3	infra-red turbidity sensors	Cal Rodó, Can Vila, Ca l'Isard
	1995-cont.	5min	2	ultrasound turbidity sensors	Cal Rodó-Ca l'Isard

<sup>1</sup>NMP= Number of measuring/sampling locations, <sup>2</sup>Meteorological variables: global and net radiation, air temperature and relative humidity, wind speed and direction, <sup>3</sup>Water isotopes: δ<sup>18</sup>O and δ<sup>2</sup>H.

#### 4. What are we learning from the Vallcebre Research Catchments?

The main results obtained since 1988 on rainfall and weather characteristics, ecohydrological processes as rainfall partitioning and forest transpiration, soil water content and piezometric level dynamics, hydrological response and runoff generation processes, as well as erosion and sediment transport, are summarized in the following sections.

##### 4.1. Rainfall, weather conditions and seasonality

Climate can be defined as humid Mediterranean, characterized by a marked water deficit in summer. Long-term (1988-2013) mean annual precipitation was  $880 \pm 200$  mm, with a mean 90 rainy days per year (Latron *et al.*, 2009). Snowfall accounts for less than 5% volume over the period. The annual precipitation at Vallcebre is characterized by strong inter-annual variability, a range of variation of more than 700 mm between extreme values and a mean difference between consecutive years of more than 200 mm.

The rainiest seasons are autumn and spring, with mean monthly rainfall amounts above 100 mm in October, November and May. Winter is the season with least precipitation. In summer, short convective storms may also provide significant precipitation input (Fig. 4 b). The spatial variability of rainfall within the catchment is limited, except during summer storms (Latron *et al.*, 2009, 2010a).

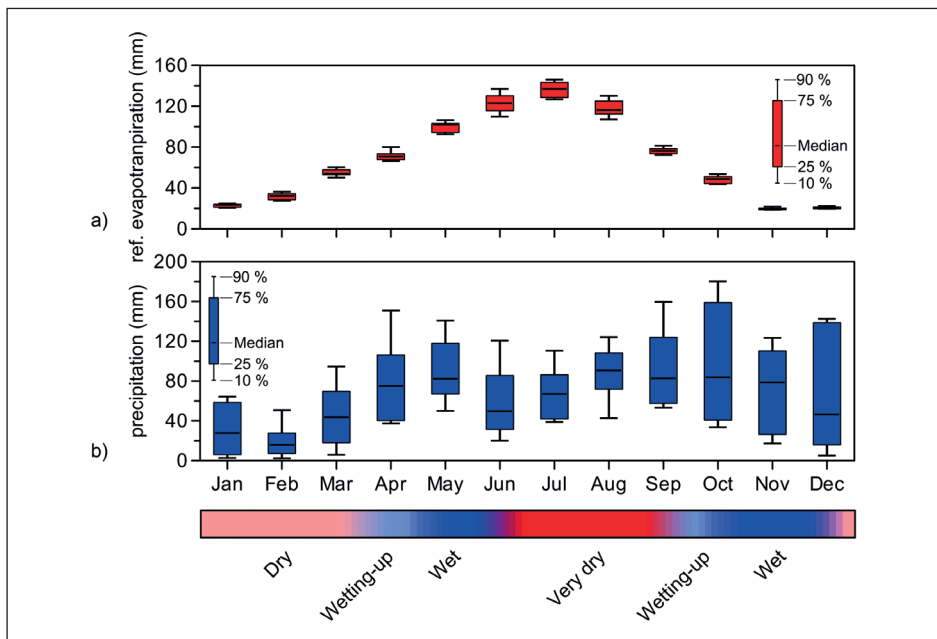


Figure 4. Annual distribution of monthly reference evapotranspiration (a) and precipitation (b) in the Vallcebre research catchments. The dynamics of rainfall and evaporative demand during the year cause the succession of dry and wet periods separated by wetting-up phases.

In more than 60% of rainfall events, fewer than 6 mm fall; in 10%, more than 25 mm; and in only 3%, more than 50 mm. Nonetheless, on average, one or two rainfall events with over 100 mm are observed each year in the area. More than 80% of the rainfall events last fewer than 6 hours; summer is usually the season with shorter and more intense rainfall events. Mean rainfall intensities are generally lower than 6 mm h<sup>-1</sup> except during the summer period. Maximum rainfall intensity in 20 minutes is less than 10 mm h<sup>-1</sup> for 80% of the events. However, 7% of the rainfall events have a maximum rainfall intensity higher than 20 mm h<sup>-1</sup>. These events are most often summer storms with high intensities that may reach 80 mm h<sup>-1</sup> in 20 minutes.

The isotopic signatures of meteoric waters in Vallcebre are influenced by the Western Mediterranean and Atlantic Ocean. The isotopic signature of rainfall in Vallcebre (2011-2016) was characterized by mean rainfall  $\delta^2\text{H}$  of  $-44.14 \pm 25.67\text{‰}$ ,  $\delta^{18}\text{O}$  of  $-7.17 \pm 3.20\text{‰}$  and deuterium excess of  $13.19 \pm 5.79\text{‰}$ . The intra-annual pattern of rainfall  $\delta^{18}\text{O}$  followed the mean monthly air temperature, but no amount effect was observed (Casellas, 2017). At the event scale, most of the episodes selected had a V-shaped isotopic trend, followed by an L-shaped one. Very few episodes showed a constant trend.

Mean annual temperature at 1260 m a.s.l. is 9.1°C and long-term (1989-2006) mean annual reference evapotranspiration, calculated by the Hargreaves-Samani (1982) method, is  $818 \pm 30$  mm. The reference evapotranspiration follows the characteristic yearly distribution (Fig. 4a), with low evaporative demand in winter, about 20 mm in December and January due to low temperatures, and high evaporative demand in summer, with a monthly maximum in July (up to 150 mm). The combination of the seasonal rainfall distribution and the evaporative demand causes a succession of dry and wet periods during the year, separated by wetting-up phases (Fig. 4).

#### 4.2. Rainfall partitioning

Rainfall partitioning for the two species studied, Scots pine and pubescent oak, indicates that throughfall is the dominant flux. Since mean throughfall and stemflow in Scots pine ran at 75% and less than 1.5%, respectively, the interception rate was 24% of the incident precipitation. For oak, mean throughfall, including the periods with and without leaves, represented 83%; and mean stemflow was slightly higher than 2%. In this case the interception losses ran at about 15% of the incident precipitation (Fig. 5).

No seasonal control of rainfall interception was found for Scots pine (Gallart *et al.*, 2002a). For this species, relative interception was similar throughout the year, due to the compensation between the characteristics of the events and the atmospheric conditions (Llorens *et al.*, 1997). For oak the differences observed between the leafed and the leafless periods were less than expected. Thus, throughfall was about 3% lower and stemflow in the leafed period was reduced to half that in the leafless period. These differences resulted in a relative interception 5% higher in the leafed period (Muzylo *et al.*, 2012a). This small difference is due to a different combination of factors prevailing in each period. These factors are related to the differences in canopy storage capacity, rainfall intensity and wind speed during the two periods. In addition, the drying times

of oak canopy are twice as long during the leaf period due to less ventilation inside the canopy (Llorens *et al.*, 2014).

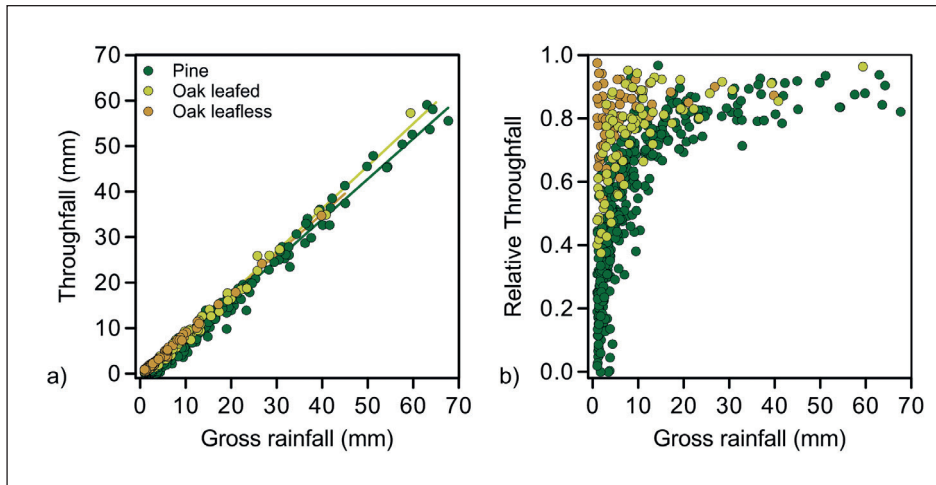


Figure 5. Throughfall as a function of incident precipitation for Scots pines and pubescent oaks leafed and leafless periods (a). Relative throughfall (throughfall divided by rainfall) as a function of incident precipitation for Scots pine and Downy oaks leafed and leafless periods (b).

At the inter-event scale short events with low rainfall intensities provide the highest interception rates, mainly due to their smaller rainfall volumes, whereas high-intensity long events have the lowest interception rates (Llorens *et al.*, 1997; Muzylo *et al.*, 2012a). The high inter-event variability of interception losses in Mediterranean conditions, described by several authors (Llorens *et al.*, 2011), can be attributed to the characteristics of the Mediterranean precipitation regime. Low-intensity events, which cause significant interception losses, are characteristic of frontal rains (e.g. Llorens *et al.*, 1997; Limousin *et al.*, 2008), while high-intensity events, which cause low interception rates, are promoted by convective storms (e.g. Llorens *et al.*, 1997; David *et al.*, 2005).

Results of rainfall partitioning modelling indicate that the original analytical Gash (1979) model, applied to the pine stand (Llorens, 1997), and the modified disperse Gash (Valente *et al.*, 1997) and Rutter (Gash *et al.*, 1995) models, applied to the oak stand (Muzylo *et al.*, 2012b) are sufficiently robust to estimate interception losses in Mediterranean conditions. Likewise, the errors of the predictions are similar to those obtained in other climates or for other species (Muzylo *et al.*, 2009). However, when model results were evaluated at the event scale, prediction errors were greater. The main reason is that, with the Gash model, a constant ratio between evaporation and rainfall rate does not allow correct simulation of the great variability of rainfall events observed in Mediterranean conditions (Limousin *et al.*, 2008; Llorens, 1997).

A detailed study of the stemflow dynamics of pine and oak (Cayuela *et al.*, in review a) indicated that differences in this flux were controlled by rainfall amount and intensity. In addition, trees' biometric characteristics were the factors differentiating the response of each tree to the same rainfall event. In fact, this study emphasizes the need of combining both biotic and abiotic factors to understand stemflow dynamics. Moreover, despite the underestimation of stemflow in most hydrological studies because it represents only a small percentage of incident rainfall (Levia and Frost, 2003; Levia and Germer, 2015), stemflow intensities, much greater for some moments than the intensity of rainfall, highlight the paramount role of stemflow as a hotspot of infiltration and soil water content dynamics.

Recent work at the plots studied on the modification of the isotopic signature ( $\delta^{18}\text{O}$  and  $\delta^2\text{H}$ ) of rainfall when passing through the canopies showed a shift in the isotopic composition of throughfall and stemflow with respect to open rainfall. In general, throughfall and stemflow tended to be more enriched in  $\delta^{18}\text{O}$  than open rainfall did, although some cases of depletion were found (Cayuela *et al.*, in review b). The modification of rainfall isotopic signature by throughfall and stemflow is attributed to the interaction of three factors: evaporation, isotopic exchange between vapour and liquid and selective canopy storage (Allen *et al.*, 2017). The analysis of these processes that take place in the canopy is of paramount importance when tracing the source and movement of water through forested or partially forested catchments (Klaus and McDonnell, 2013).

#### 4.3. Forest transpiration and ecohydrological processes

Transpiration of the Scots pine was double that of the nearby pubescent oak (Poyatos *et al.*, 2005), largely due to differences in the leaf area index between the stands. Maximum transpiration rates at optimal soil water content conditions were higher in the Scots pine stand than in the pubescent oak stand. However, when soil water content became a limiting factor, transpiration in the Scots pines dropped considerably, whereas the pubescent oaks were less affected by soil water-content deficits. The physiological responses to water deficits of both species indicated that Scots pine was more vulnerable to xylem embolism, whereas pubescent oak was more resistant to extreme drought events (Poyatos *et al.*, 2008). There was a 40% greater reduction of Scots pine total summer transpiration during dry summers than in a summer with average rainfall. Moreover, full transpiration recovery after significant rainfall events was not observed during dry summers, which could indicate some lasting drought effects, like the embolism mentioned above (Llorens *et al.*, 2010). Likewise, the relationship found between climate characteristics and growth indicates a strong effect of past growth trends and current water use strategies on tree resilience to increased aridity, which is more evident in Scots pine (Morán-López *et al.*, 2014). The large proportion of rainfall used for Scots pine transpiration during dry summers results in a strong reduction of runoff and deep water stores at the catchment scale, suggesting that the predicted increase in the frequency of severe summer droughts may threaten the current role of Mediterranean mountain catchments as suppliers of water resources for lowland areas (Llorens *et al.*, 2010).

Sapflow data measured at the pine plot have been included in several meta-analyses of *Pinus sylvestris*' physiological adjustment to climatic conditions, growth and resistance to drought (Poyatos *et al.*, 2007; Martínez-Vilalta *et al.*, 2009; Sterck *et al.*, 2012). Likewise, sapflow data, combined with rainfall, meteorological and soil water content data, have been used to validate remote sensing (Cristobal *et al.*, 2011) and modelling approaches (De Cáceres *et al.*, 2015; Sus *et al.*, 2014).

Recently, a dual isotope-based approach ( $\delta^{18}\text{O}$  and  $\delta^2\text{H}$ ) has been used to examine the mixing of water in the soil and the linkages between tree water fluxes and soil water pools. Preliminary results indicate that stable isotope ratios of bound soil water fell below the local meteoric water line (LMWL), with more evaporative enrichment in the shallow horizons. On the contrary, mobile soil water (low suction cup lysimeters) and groundwater fell along the LMWL, well mixed with stream water.

#### 4.4. Soil water content and piezometric level dynamics

At the catchment scale, soil water content shows a seasonal pattern characterized by the occurrence of periods of marked deficit in summer and, though less pronounced, in winter (Fig. 6a). The spatial variability of soil water content (0-80 cm depth profiles) was not consistent throughout the year. Differences were minimal during the driest and wettest conditions, whereas the greatest differences between locations were observed under intermediate moisture conditions (Fig. 6b), during the drying and wetting-up periods (García-Estringana *et al.*, 2013). Soil water content during wetting-up and drying-down periods probably depended mainly on local controls (Grayson *et al.*, 1997), such as vegetation and soil properties. In fact, results indicated (Fig. 6c) that soil water content on hillslopes under forest cover was lower than in downslope areas covered by grasses (Gallart *et al.*, 1997, 2002a). However, during wet conditions, soil water content patterns were controlled by non-local factors. Under these conditions, the general topographic features of the catchment, favouring water redistribution, limited the effect of the vegetation on soil water content spatial patterns (García-Estringana *et al.*, 2013).

Results of soil water content spatial variability monitored on one abandoned agricultural terrace showed that, at the daily scale, variability increased with soil water content. Moreover, as observed at the catchment scale, a low effect of local controls under dry conditions and a higher effect of non-local controls under wetter conditions were observed. At this spatial scale the main non-local control observed on soil water content was the distance of each location from the inner part of the terrace, because the inner parts of the terraces are areas prone to saturation due mainly to shallow soil depth (Molina *et al.*, 2014). With this information the representativeness of different spatial and temporal designs were tested. Results indicated that 10 x 10 m grids were a robust design for obtaining a good estimation of mean soil water content dynamics. During wetting-up periods the longest time-step that could detect rapid changes in soil water content was 8 h, whereas during drying-down periods soil water content monitoring frequency of 24 h was enough to catch changes in water balance (Molina *et al.*, 2014).

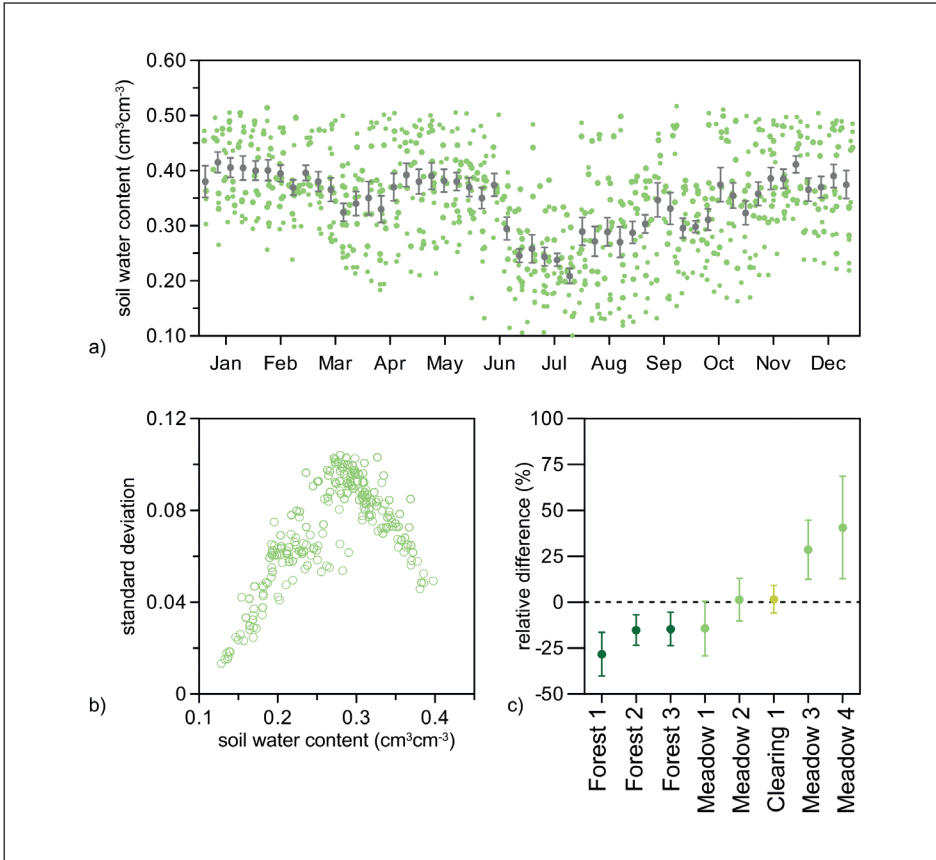


Figure 6. Intra-annual evolution (weekly measurements from 1994 to 2003) of mean soil water content at a mid-slope location, in the Cal Parisa catchment (a). Grey dots correspond to mean values and standard errors obtained from all available measurements for a given week. Relationship between mean soil water content (averaged from different locations in the Cal Parisa catchment) and its standard deviation as an indicator of soil water content spatial variability (b). Ranked values (from dry to wet) of the relative difference between the mean soil water content value of each location and mean value of all the locations in the Cal Parisa catchment (c). Error bars =  $\pm 1$  standard deviation. DOY: Day Of the Year.

Shallow groundwater dynamics also show a seasonal pattern at the catchment scale. During the summer period, the water table drops markedly every year in response to increasing evapotranspiration (Fig 7a). As for soil water content, the water table rises steadily during the wetting-up period in response to autumn rainfall events and often remains in the first 50 cm below the surface during winter and spring.

The analysis of shallow groundwater dynamics during rainfall-runoff events indicated marked spatial variability, controlled mainly by the distance of the measuring location (piezometer) from the stream (Fig. 7b). The analysis of the effect

of antecedent wetness conditions on runoff generation distinguished three processes. In dry conditions, with a marked spatial variability of piezometric levels, shallow groundwater located near the stream channel was the main contributor to storm-flow. In intermediate conditions, there was a more generalized response of shallow groundwater to rainfall, though the timing between different locations was variable. Finally, in wet conditions a pre-event water table close to the soil surface gave more homogeneous responses (Latron and Gallart, 2008; Latron *et al.*, 2010; Roig-Planasdemunt *et al.*, submitted).

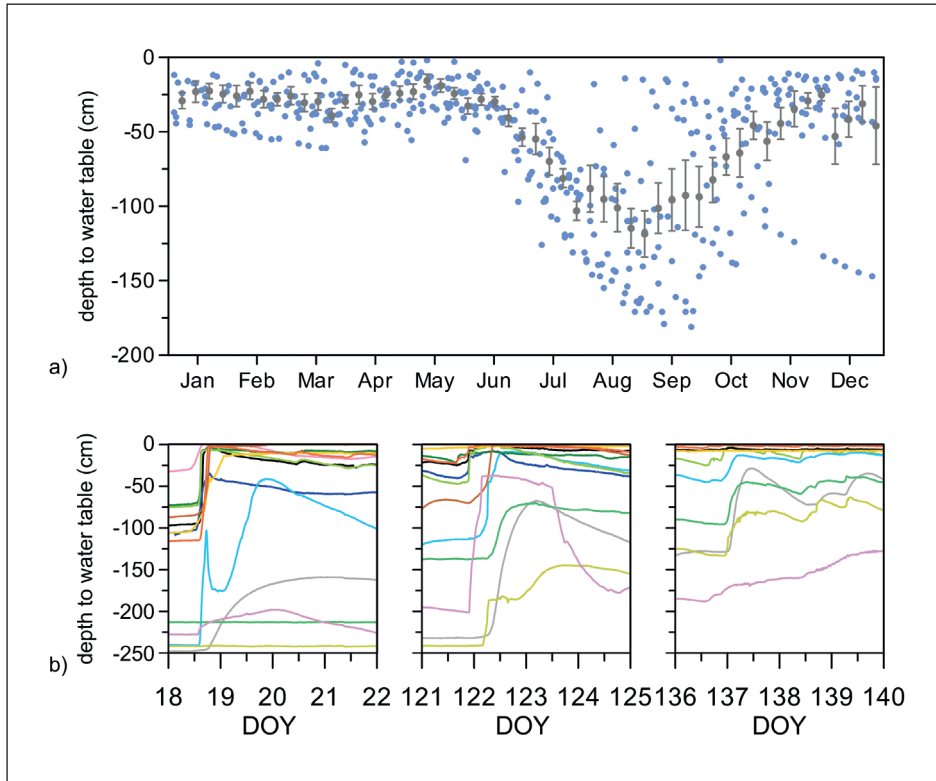


Figure 7. Intra-annual evolution (weekly measurements from 1994 to 2003) of water table depth, in the Cal Parisa catchment. Grey dots correspond to mean values and standard errors obtained from all available measurements for a given week (a). Depth to water table dynamics observed at 13 piezometer locations in the Can Vila catchment, during floods occurring in dry (left), wetting-up (middle) and wet (right) conditions (b). DOY: Day Of the Year.

#### 4.5. Hydrological response and runoff generation processes

During the water years 1995-2012, mean annual runoff in the Cal Rodó catchment was  $210 \pm 121$  mm (equivalent to 25 % of rainfall) and in Can Vila it was  $305 \pm 196$  mm (36% of rainfall). In spite of this difference in annual runoff, related to the deep percolation in



the limestone area of the Cal Rodó catchment, the relationship between annual runoff in both catchments was broadly linear. Mean daily discharge over the period was  $6.7 \text{ ls}^{-1} \text{ km}^{-2}$  (min=0; max= $1093 \text{ ls}^{-1} \text{ km}^{-2}$ ) in Cal Rodó and  $9.7 \text{ ls}^{-1} \text{ km}^{-2}$  in Can Vila (min=0; max= $1024 \text{ ls}^{-1} \text{ km}^{-2}$ ). However, stream-flow was highly seasonal, and the streams dried out in summer every 2 years on average, for a period ranging from 15 to 40 days. Stream-flow is rather flashy in the Cal Rodó and Can Vila catchments, with response time of around one hour. Recessions are quite steep in the Cal Rodó catchment, but gentler in Can Vila.

There was no simple rainfall-runoff relationship on either seasonal or monthly scales. Monthly analysis revealed the existence of a threshold in the rainfall-runoff relationship, mainly due to the non-linearity observed between groundwater table and runoff, as indicated above (Latron *et al.*, 2008). At the event scale, the storm-flow coefficient showed a clear seasonal pattern. During wet periods, at the end of autumn or beginning of winter, the highest storm-flow values were observed. In these conditions the catchment is hydrologically active and shows a significant response to precipitation. On the contrary, the lowest storm-flow values were observed during summer. Moreover, in dry conditions storm-flow coefficient variability was frequently two orders of magnitude larger than during wet conditions (Gallart *et al.*, 1997; Latron *et al.*, 1997, 2010). The relationships between the storm-flow coefficient and rainfall depth, rainfall maximum intensity and baseflow showed that the correlations found were the same as those observed for wet climates, even if correlation coefficients were notably lower (Latron *et al.*, 2008).

The combined analysis of soil water tension, groundwater table dynamics and runoff contributing areas during floods identified three types of hydrological behaviour (Latron and Gallart, 2007, 2008) (Fig. 8). In dry conditions, only limited storm-flow response was observed at the catchment outlet due to, as indicated above, low groundwater tables and dry soils. Under these conditions, catchment response was fast and short: neither topsoil saturation nor water table rise were detected and thus no saturated area was observed (Latron and Gallart, 2007). For these events the only possible runoff-contributing areas were local, low-permeable areas (bedrock outcrops or badlands) close to the stream, with infiltration excess runoff probably the dominant runoff process. During transitions from dry to wet conditions, as a response to more frequent and/or greater rainfalls, larger runoff events with moderate storm-flow coefficients were observed. These events were characterised by a long response time and a relatively gentle recession. A scattered saturation pattern, due to a perched saturation layer reaching the surface at some places favoured by the terraced topography and soil characteristics, was observed during these conditions (Latron and Gallart, 2007). Saturation excess overland flow on perched saturated layers was probably the dominant runoff generation process during the wetting-up periods. In wet conditions, when the catchment was saturated or close to saturation, large and lasting runoff response was observed. In these conditions, the shallow groundwater table rose quickly, causing saturated areas to expand rapidly. Excess overland flow on saturated areas was most probably the main runoff generation process in these conditions.

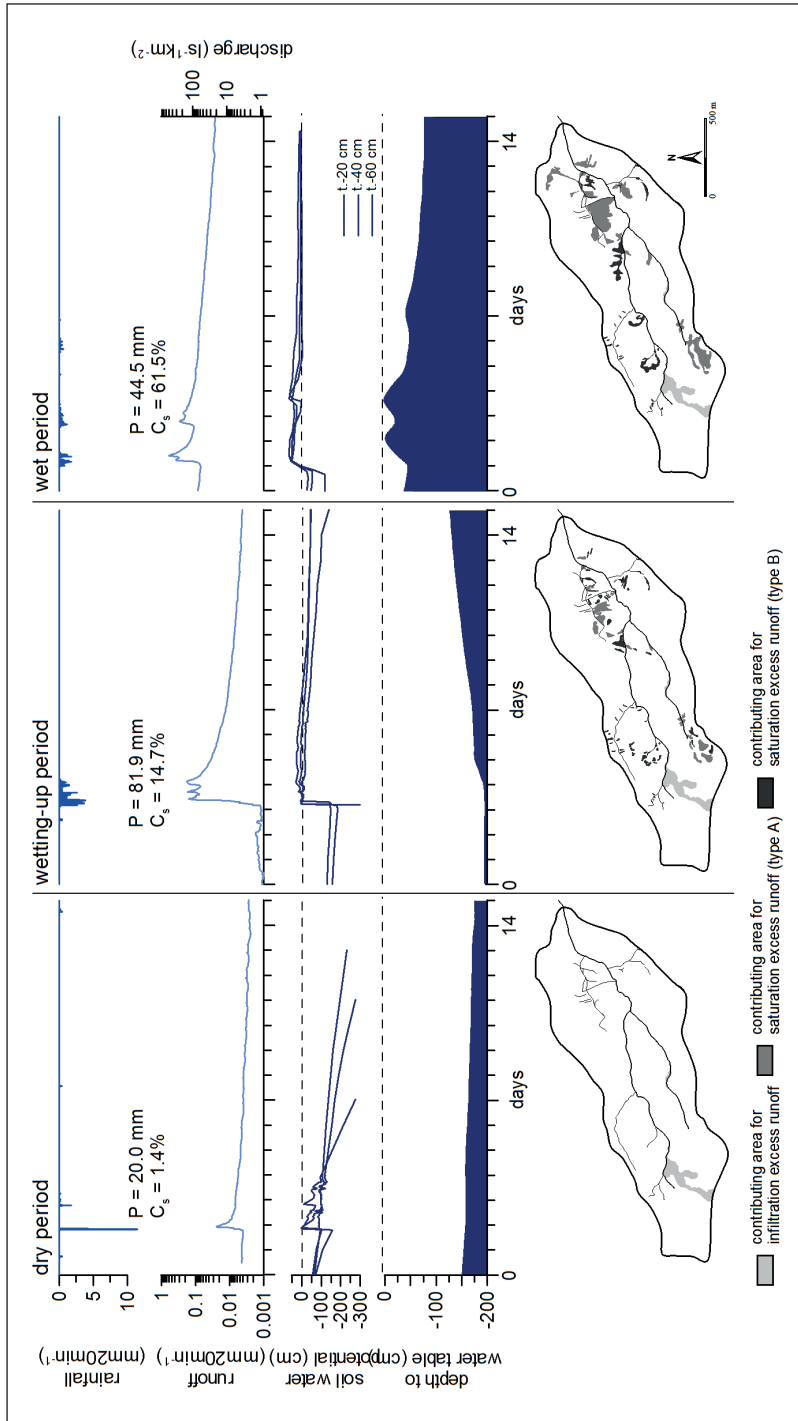


Figure 8. Soil water potential and water table depth dynamics and spatial pattern of contributing areas observed for characteristic rainfall runoff events during dry, wetting-up and wet conditions. ( $P$  = rainfall depth;  $C_s$  = storm-flow coefficient). Type A-saturated areas are located at flow convergence sites and result from groundwater rise to the surface. Type B-saturated areas are linked to local soil properties allowing the frequent development of a perched saturated layer.

#### 4.6. Environmental tracers and transit time

The study of dissolved organic carbon (DOC) dynamics in the Can Vila catchment showed some seasonality in precipitation and soil water, both related to biological activity (Roig-Planasdemunt *et al.*, 2017). However, no clear seasonality was observed in groundwater, where DOC and water table dynamics were closely related. On the other hand, DOC in stream water was clearly dependent on discharge, but there was discontinuity, which depended on the magnitude of the discharge. Low and constant DOC concentrations were observed for low flow conditions ( $<20 \text{ l s}^{-1} \text{ km}^{-2}$ ), whereas DOC concentration increased as discharge increased during flood conditions. The increase in stream water DOC concentration during floods suggests a relevant contribution of soil water. Moreover, a DOC flushing effect observed at the first discharge peak, in events with several peaks, suggests the presence of DOC sources in the streambed. However, similar stream water DOC dynamics were observed during storm events with diverse antecedent conditions and event characteristics, raising the question of the origin of rapid DOC increase during floods.

Two-component hydrograph separation results, using stable isotopes in the Can Vila catchment, indicate that pre-existing water contributed between 30% and almost 100% to total runoff (Fig. 9), depending on antecedent moisture conditions, the extent of saturated areas within the catchment and precipitation characteristics (Latron *et al.*, 2015). During low to moderate intensity rainfall events (winter, spring, autumn), old water contribution was dominant along the hydrographs; new water contribution was always less than 10%. Conversely, during high-intensity summer storms, old water contribution in the hydrograph was lower (30% to 50%) and the response corresponded mostly to new water. A negative correlation was found between maximum rainfall intensity and old water contribution at the event scale. In addition, the lower the old water contribution in the hydrograph, the higher the maximum suspended sediment concentration observed at the outlet, which suggests that summer floods were mainly caused by infiltration excess runoff generated on badlands and degraded areas of the catchment (Latron *et al.*, 2015).

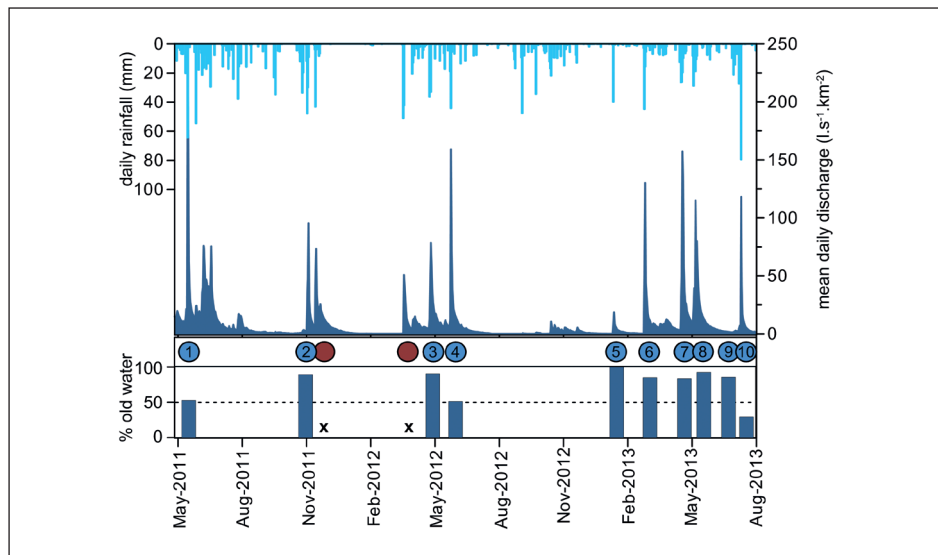


Figure 9. Time series of daily rainfall and mean daily discharge measured in the Can Vila catchment (May 2011–July 2013). Numbers from 1 to 10 refer to floods sampled for stable isotopes. Old water contribution derived from two-component hydrograph separation ( $^{18}\text{O}$ ) is shown for each flood. Red circles correspond to two floods for which hydrograph separation was not possible because of similar isotopic signals in rainfall and in the stream before the flood.

A new methodology to assess the mean transit time (MTT) and its uncertainty was recently developed and applied to the Vallcebre research catchments (Gallart *et al.*, 2016). This methodology, using a general likelihood uncertainty assessment (GLUE), takes into account both the analytical and model parameter errors. Using a resampling procedure, it determines the statistical significance of the MTT differences found between pairs of sampled sites. This methodology was applied to Tritium samples taken during steady-state baseflow periods in diverse streams, wells and springs in the Vallcebre research catchments in the late nineties (1996–1998) by Hermann *et al.* (1999) and in 2013. Results indicate that water samples can be classified in three main groups, consistent with the geological setting of the catchment: shallow open aquifers with MTT of nearly 5 yrs; spring and stream waters from one subcatchment with MTT of nearly 7.5 yrs; and from another subcatchment, with MTT of nearly 12 yrs. These results support the use of tritium in MTT studies to gain information about old waters not identifiable by stable isotopes (Gallart *et al.*, 2016).

#### 4.7. Erosion and sediment transport processes

Badlands are the main sediment sources in the Vallcebre research catchments. To understand the erosion and sediment dynamics of these areas, two main approaches have been used (see Gallart *et al.*, 2013 for a review). On the one hand, short-term studies calculated denudation rates and sediment production at the plot

scale, to analyse weathering seasonal dynamics and to study the relationships between geomorphic activity and herbaceous plant colonisation. On the other hand, long-term studies (1990 till now) studied suspended sediment dynamics at the catchment scale. Results indicate that ground freezing in winter is the main weathering mechanism, whereas erosion is especially active in summer due to both the availability of erodible material and the occurrence of intense rain events. However, the eroded sediments may remain deposited in the streams from months to years (Gallart *et al.*, 2013).

Results on denudation rates and sediment production at the plot scale gave rates between 11,000 and 23,000 Mg km<sup>-2</sup> yr<sup>-1</sup>, depending on the method used. The uncertainty of these estimates was analysed by applying the KINEROS2 model to a set of badland areas for a period of 15 years, giving a long-term sediment production of about 9000 Mg km<sup>-2</sup> yr<sup>-1</sup>. This also allowed a detailed analysis of the temporal patterns of the erosive events (Martínez-Carreras *et al.*, 2007). Results at the catchment scale, recorded for 18 years at the Ca l'Isard sub-basin (with 4.5% of the surface covered by badlands), showed a sediment yield of 534 Mg km<sup>-2</sup> yr<sup>-1</sup>. In addition, this long-term record allowed us to analyse the uncertainty of the measurements associated with the length of the recording period. The combination of the three approaches, plot and catchment scales and modelling, demonstrated that records longer than 5 years are needed to assess the order of magnitude of the erosion or transport rates (Gallart *et al.*, 2013).

Recently, the assessment of sediment connectivity was analysed by use of lead-210 (<sup>210</sup>Pb<sub>ex</sub>) fallout in the Cal Rodó catchment. Results indicated low sediment <sup>210</sup>Pb<sub>ex</sub> concentrations in both the sediments deposited in the streambed and the suspended sediments. These low concentrations are the result of the fresh-rock origin of sediments eroded in badland areas. Results indicated that fine suspended sediment <sup>210</sup>Pb<sub>ex</sub> activity follows the seasonal pattern of fallout accumulation during most of the year and erosion by summer storms, and that <sup>210</sup>Pb<sub>ex</sub> activity in fine suspended sediments decreased towards the catchment outlet. These results suggest high connectivity between the badlands, streams and catchment outlet for fine sediments, but much lower connectivity of sandy or coarser sediments eroded from badland surfaces and untagged by the fallout, as they mainly remain in the stream bed (Moreno de las Heras *et al.*, 2017).

Badlands at Vallcebre, and more generally in the high Llobregat catchment, are examples of humid badlands (Gallart *et al.*, 2002b, 2013). In these intensely eroded areas, vegetation is not limited by dryness (as in arid or semi-arid badlands), but by the severe weathering and erosion processes driven by freezing-thawing processes and summer convective storms, respectively (Regüés *et al.*, 2000). Vegetation is more abundant and diverse on sunny (south-facing) aspects than on shady ones because low winter temperature is a more limiting factor than summer drought for Mediterranean species (Guàrdia *et al.*, 2000). Badland surfaces develop on north-facing aspects or those with no clear orientation, depending on the higher or lower strength of bedrock (Moreno de las Heras and Gallart, 2016).

## 5. Conclusions

Thirty years of research in the Vallcebre Research Catchments have increased understanding not only of their hydrological response, but also of the main hydrological and erosion processes characteristic of Mediterranean mountain catchments.

The main findings from this research are the following:

- The marked temporal variability in precipitation, both intra-annual and seasonal, clearly determines the hydrological processes.
- The high inter-event variability of interception losses can be attributed to the characteristics of the Mediterranean precipitation regime, which includes frontal rains and convective storms, producing high and low interception losses, respectively.
- There is a general tendency for throughfall and stemflow to be more  $\delta^{18}\text{O}$ -enriched than bulk rainfall.
- Scots Pine is more vulnerable to drought than pubescent oak. Moreover, the large proportion of rainfall used for Scots pine transpiration during dry summers results in a strong reduction of runoff and water stores.
- Forest cover, as against meadows, plays a paramount role in reducing soil water content during intermediate moisture conditions.
- There is a low effect of local controls on soil water content under dry conditions and a higher effect of non-local control under wetter conditions, at catchment and plot scales.
- There is a marked influence on runoff generation of antecedent wetness conditions, soil water content and shallow groundwater.
- Complex rainfall-runoff relationships at seasonal and monthly scales were observed, determined by the non-linearity between groundwater and runoff responses.
- There were three types of hydrological behaviour: areas contributing partially during dry periods, a wetting-up transition period characteristic of Mediterranean catchments and behaviour similar to those reported for humid climates in wet conditions.
- During floods, the contribution of pre-existing water is often dominant.
- There is no clear indication of DOC origin during the different type of floods.
- Mean transit time determination supports the use of tritium to identify old waters, not identifiable by stable isotopes.
- Records longer than 5 years are needed to assess the magnitude of erosion or transport rates.
- Badlands at Vallcebre provide examples of humid badlands, where vegetation is controlled by severe weathering, and erosion driven by freezing-thawing processes and summer convective storms.

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